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## Exposure to per- and polyfluoroalkyl substances (PFAS) and type 2 diabetes risk

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### Abstract

Per- and polyfluoroalkyl substances (PFAS) are ubiquitous man-made chemicals found in consumer products including fabrics, food packaging, non-stick coatings, and aqueous film-forming foams. PFAS are stable and extremely resistant to degradation, resulting in high persistence throughout the environment as well as in human blood. PFAS consist of a large family of synthetic chemicals, with over 4000 distinct varieties having been identified and around 250 currently being manufactured at globally relevant levels. Numerous epidemiological studies have linked exposure to PFAS with adverse health effects ranging from immunotoxicity, cardiometabolic disease, developmental and reproductive effects, cancer, and recently type 2 diabetes. Several studies have demonstrated associations between serum PFAS concentrations and glycemic indicators of type 2 diabetes including glucose, insulin, and HOMA-IR in adolescent and adult cohorts. In addition, some studies have shown positive associations with incident type 2 diabetes and multiple PFAS. However, the link between PFAS exposure and the development of diabetes continues to be a disputed area of study, with conflicting data having been reported from various epidemiological studies. In this mini review we will summarize the current state of the literature linking PFAS to type 2 diabetes and discuss important future directions including the use of more complex mixtures-based statistical analyses.



**Keywords:** PFAS toxicity, Per- and polyfluoroalkyl substances, diabetes, cardiometabolic disease, hyperglycemia, insulin

### Introduction

Diabetes has set a high burden on both individuals and health care systems, with global health expenditure due to diabetes in 2021 reaching over \$966 billion (1). Diabetes is a chronic, metabolic condition in which individuals have elevated blood glucose levels, also known as hyperglycemia. When left untreated, hyperglycemia can lead to serious damage to blood vessels, nerves, and various tissues and organs, especially the heart and kidneys (2). In the United States, diabetes is the seventh leading cause of death (3). Much of this is due to diabetes being a major risk factor for the development of cardiovascular disease (CVD), which is the most common cause of death in those with Type 2 diabetes (T2D), accounting for over 52% (4, 5). Diabetes occurs when the body is not able to produce enough insulin (T1D) or when the body cannot effectively use the insulin that is being produced (T2D) (6). Increases in blood glucose leads to increased synthesis and secretion of insulin by beta cells in the pancreas. Insulin is an endocrine peptide hormone that binds to receptors in target cells and activates signaling pathways resulting in various metabolic changes, most notably increased glucose uptake (7). T2D is the most common type of diabetes and occurs when your pancreas does not produce enough insulin and cells become resistant to insulin, leading to lower cellular uptake of glucose and therefore higher blood glucose levels (3, 6). Diabetes is commonly diagnosed by measuring glucose and/or insulin after an oral glucose tolerance test or by hemoglobin A1C test (8). Insulin resistance, another hallmark of diabetes, can be measured by calculating the homeostatic model assessment (HOMA-IR) (9). Long-standing studies have demonstrated how genetic and lifestyle factors have a major impact on the development and progression of T2D (10–12). Several genes involved in beta-cell development or function have been identified through genome-wide association studies that confer T2D risk (11, 12). In addition, little physical activity, a poor diet, smoking, stress, and high body mass index have all been attributed in the promotion of T2D prevalence and incidence (10). However, studies have begun to investigate additional environmental factors that potentially impact on the development of diabetes, such as exposure to environmental contaminants acting as endocrine-disrupting chemicals, including per- and polyfluoroalkyl substances (13–15).

Per- and polyfluoroalkyl substances (PFAS) are man-made chemicals containing a fluorinated carbon chain that act as persistent organic pollutants (POPs) (16). These chemicals have been manufactured and utilized for decades in countless consumer and industrial products for their excellent surfactant properties (17, 18), including fabrics, food packaging, non-stick coatings, and aqueous film-forming foams (16, 19–21). PFAS as a class of chemicals may represent upwards of 4000 distinct varieties of which 250 are currently being manufactured at globally relevant levels (22, 23). PFAS are very stable and extremely resistant to degradation, resulting in high persistence throughout the environment, including water sources, soil, plants, animals, as well as humans around the globe (24–27). Humans are exposed to PFAS *via* several exposure routes, the principal ones being through contaminated drinking water or food supplies (26, 28). Even among the general population, PFAS can be detected in the blood, with data from the National Health and Nutrition Examination Survey (NHANES) 2003–2004 reporting that PFAS can be measured in 98% of U.S. adults (29–31). PFOS and PFOA were measured in serum at 20.7 ug/L and 3.9 ug/L averaged for the whole population, respectively (29). Although legacy halogenated POPs, such as PCBs, have

## PFAS and glucose homeostasis

Since PFAS are persistent organic pollutants ubiquitous throughout the environment, bioaccumulating in humans across the globe, research exploring the adverse health effects associated with exposure to PFAS has greatly increased over the past few decades (34, 35). PFAS have been associated with multiple chronic diseases in epidemiological studies including immunotoxicity, cardiometabolic disease, developmental and reproductive effects, and cancer (36–40). More recent studies have begun to explore associations between PFAS exposure and diabetes (14, 15). However, the link between PFAS exposure and the development of diabetes continues to be a disputed area of study, with conflicting data having been reported from various epidemiological studies. **Table 1** outlines many of these epidemiological studies examining the association of PFAS exposure with the development of T2D. Of these, several studies have demonstrated positive associations between serum PFAS concentrations and an increase in glycemic indicators of T2D. In 2016, Domazet et al. analyzed data from the European Youth Heart Study, which collected samples from participants from 1997-2009 during childhood (9 years), adolescence (15 years) and young adulthood (21 years) (41). PFOA exposure in childhood was found to be associated with decreased  $\beta$ -cell function at 15 years of age (HOMA- $\beta$ :  $\sim$  -12% decrease change in outcome by 10 ng/ml increase in PFOA exposure). In 2016, Kim et al. published results from a double-blind, placebo-controlled study with elderly (60+ years) Korean subjects recruited in 2011 that were given vitamin C treatments for 4 weeks (42). Of the 8 PFAS measured, the authors found PFOS and PFDoDA to be positively associated with HOMA-IR at baseline and after placebo treatment ( $\beta_{\text{PFOS}} = 0.60$  increase by IQR; 95% CI: -0.002, 1.19,  $p=0.053$ ;  $\beta_{\text{PFDoDA}} = 0.83$ ; 95 % CI 0.23, 1.44,  $p=0.01$ ). Interestingly, these effects were not observed after vitamin C treatment. In 2017, Cardenas et al. examined associations between plasma PFAS concentrations and glycemic indicators from participants of the Diabetes Prevention Program (DPP), who were recruited between 1996-99 (43). At baseline, a doubling in PFOS and PFOA concentrations was associated with increased HOMA-IR ( $\beta_{\text{PFOS}} = 0.39$ ; 95% CI: 0.13, 0.66;  $\beta_{\text{PFOA}} = 0.64$ ; 95% CI: 0.34, 0.94], increased HOMA- $\beta$  ( $\beta_{\text{PFOS}} = 9.62$ ; 95% CI: 1.55, 17.70;  $\beta_{\text{PFOA}} = 15.93$ ; 95% CI: 6.78, 25.08), increased fasting proinsulin ( $\beta_{\text{PFOS}} = 1.37$  pM; 95% CI: 0.50, 2.25;  $\beta_{\text{PFOA}} = 1.71$  pM; 95% CI: 0.72, 2.71), and increased HbA1c ( $\beta_{\text{PFOS}} = 0.03\%$ ; 95% CI: 0.002, 0.07;  $\beta_{\text{PFOA}} = 0.04\%$ ; 95% CI: 0.001, 0.07). However, no strong associations were found during the follow-up period for PFAS and the above diabetes indicators. In 2019, a study was published that analyzed serum samples from a cohort in Tianjin, China ranging in age from 19 to 87 years old that had been recruited in 2017 (44). The authors found that a 1% increase in PFOA and PFNA had significant positive associations with fasting glucose levels: 0.018% [(95% CI): 0.004%, 0.033%] and 0.022% (95% CI: 0.007%, 0.037%), respectively. In addition, a 1% increase in PFNA, PFHxA, and PFHxS were had a significant positive association with HbA1c levels: 0.018% (95% CI: 0.003%, 0.033%), 0.030% (95% CI: 0.010%, 0.051%), 0.007% (95% CI: 0.003%, 0.011%), respectively. In 2019, Alderete et al. published an analysis of a cohort of overweight and obese Hispanic children (8-14 years) recruited from 2001-2012 from Los Angeles, California (45). A 2-hour OGTT was performed at baseline and at a follow-up 1-3 years later. The study found that each ln(ng/ml) increase in PFOA and PFHxS was significantly associated with an increase in 2-

17.7)). In 2021, Zeeshan et al. analyzed data from the “Isomers of C8 Health Project in China,” which collected samples from participants 35+ years old between 2015-2016 ([46](#)). The levels of several PFAS were measured, including PFOA, PFOS, PFHxS, PFHpS, PFNA, PFDA, PFUnDA, and PFTrDA. For isomers of PFOS, PFOA, PFHxS, PFHpS, PFNA, PFDA, and PFUnDA significant positive associations were found with fasting blood glucose, fasting insulin, and HOMA-IR. For example, an increase in branched PFOS was determined to be positively associated with fasting blood glucose levels ( $\beta_{\text{PFOS}} = 0.25$ , 95% CI: 0.18, 0.33), fasting insulin ( $\beta_{\text{PFOS}} = 2.19$ , 95% CI: 1.44, 2.93) and HOMA-IR ( $\beta_{\text{PFOS}} = 0.69$ , 95% CI: 0.50, 0.89). In 2021, Valvi et al. analyzed data from a prospective cohort made up of those born in the Faroes Islands (born 1986-87) ([47](#)). Blood was collected and PFAS measured in cord whole blood and at ages 7, 14, 22, and 28 years. PFOS exposure was associated with a decrease in insulin sensitivity and an increase in insulin secretion. For example, the log-ISI per PFOS doubling was:  $\beta_{\text{PFOS}} = 0.12$  (95%CI: 0.02, 0.22), for prenatal exposures. Although associations for PFOA and PFHxS were often in the same direction as PFOS, they were not statistically significant. In addition, the overall life-course of PFOS exposure was found to be associated with altered glucose homeostasis ( $p=0.04$ ).

Key physiological endpoints from studies that show positive associations between PFAS exposure and markers of T2D risk.

Citation	Cohort	PFAS Serum Concentrations	Physiological Endpoints
Domazet et al., 2016 ( <a href="#">41</a> )	European Youth Heart Study (1997-2009)	9-year Median PFOA (ng/mL): • Male = 9.7 • Female = 9.0	↓ Decreased HOMA-β (at 15 yrs) significantly associated with PFOA (at 9 yrs)
Kim et al., 2016 ( <a href="#">42</a> )	Vitamin C intervention study in the elderly (60+ years old) in Seoul, Korea (2011-2012)	Mean (ng/mL): • PFOS = 10.04 • PFOA = 4.61 • PFDODA = 0.25	↑ Increased HOMA-IR significantly associated with PFOS and PFDODA
Cardenas et al., 2017 ( <a href="#">43</a> )	Diabetes Prevention Program (1996-1999)	Geometric mean (ng/mL): • PFOA = 4.82 • PFOS = 26.38	↑ Increased HOMA-IR significantly associated with PFOS and PFOA ↑ Increased HbA1c significantly associated with PFOS and PFOA
Duan et al., 2020 ( <a href="#">44</a> )	Workers at Nankai University, Tianjin, China (19-87 years old) (2017)	Geometric mean (ng/mL): • PFOA = 14.82 • PFNA = 3.14 • PFHxS = 0.06 • PFHxA = 0.54	↑ Increased fasting glucose significantly associated with PFOA and PFNA ↑ Increased HbA1c significantly associated with PFNA, PFHxS, and PFHxA
Alderete et al., 2019 ( <a href="#">45</a> )	Study of Latino Adolescents at Risk of T2D (Overweight and obese Hispanic children 8-14 yrs old) (2001-12)	Geometric mean (ng/mL): • PFOA = 2.78 • PFOS = 12.22 • PFHxS = 1.65	↑ Increased 2-hour glucose levels (GTT) significantly associated with PFOA and PFHxS
Zeeshan et al., 2021 ( <a href="#">46</a> )	Isomers of C8 Health Project in China (35+ years old) (2015-2016)	Median (ng/mL): • br-PFHxS = 0.01 • n-PFOA = 4.63 • n-PFOS = 9.95 • br-PFOS = 14.25 • n-PFHxS = 0.60	↑ Increased fasting glucose significantly associated with br-PFOS, n-PFOS, br-PFHxS, n-PFHxS, n-PFOA, PFHpS, PFNA, PFDA, and PFUnDA ↑ Increased HOMA-IR significantly associated with br-PFOS, n-PFOS, br-PFHxS, n-PFHxS, n-PFOA, PFHpS, PFNA,

Down arrow = decreased.

Although several studies demonstrated a positive association between serum PFAS levels and an increase in diabetes indicators, various studies have also determined there to be a non-significant or inverse association, as outlined in [Table 2](#). In 2018, Liu et al. examined subjects from the NHANES (2013-14) to analyze associations between isomers of PFOA/PFOS and a wide-ranging biochemistry profile ([48](#)). They found that branched PFOA and linear PFOS were significantly associated with decreased fasting glucose ( $\beta_{br-PFOA} = -4.07 \pm 1.69 \text{ mg/dL (SE)}$ ) ( $\beta_{n-PFOS} = -1.8 \pm 0.72 \text{ mg/dL}$ ). No significant associations with PFAS were found for 2hr glucose (GTT), insulin levels, or HOMA-IR. All forms of PFOA were found to be associated with decreased HbA1c ( $\beta_{PFOA} = -0.12 \pm 0.05$ ). Interestingly,  $\beta$ -cell function, measured by HOMA2-IR, was found to be significantly increased by n-PFOA ( $\beta_{n-PFOA} = 0.12 \pm 0.05$ ). In 2010, Nelson et al. analyzed data from NHANES (2003-04) for associations between PFAS (PFOA, PFNA, PFOS, and PFHxS) and insulin resistance, but found no significant association between the PFAS examined and HOMA-IR ([49](#)). In 2012, Fisher et al. analyzed data from the Canadian Health Measures Survey (2007-09) for associations between PFOA, PFOS, and PFHxS levels and plasma glucose, plasma insulin, and HOMA-IR ([50](#)). No association was found between the PFAS examined and these diabetes indicators. In 2017, Fleisch et al. analyzed blood samples from children (~8 years) in Project Viva, a cohort from the Boston, Massachusetts area that had samples collected from 1999-2002 for prenatal visits and 2007-2010 for child plasma ([51](#)). Children that had higher concentrations of PFOA, PFOS, and PFDeA were associated with lower HOMA-IR per interquartile range (PFOA: -10.1% (95% CI: -17.3, -2.3), PFOS: -10.1% (-16.4, -3.3), PFDeA: -14.7% (-22.1, -6.5)). Sex-specific patterns were also observed between childhood PFAS concentrations and HOMA-IR, with childhood PFOA, PFOS, and PFDeA levels associated with decreased HOMA-IR in females but not significantly associated in males. In 2019, Fassler et al. published a cross-sectional study of young girls (6-8 years) from the Cincinnati area recruited from 2004-2006 ([52](#)). While no significant associations were found between PFAS and fasting insulin and glucose, PFOA levels were found to have a borderline inverse statistical association with fasting insulin levels ( $p=0.09$ ), insulin resistance (HOMA-IR) ( $p=0.09$ ), and a positive borderline effect on insulin sensitivity (ISI) ( $p=0.10$ ). Many of the studies demonstrating a negative or non-significant association between PFAS exposure and glucose/insulin levels and resistance analyzed national health surveys with lower risk populations and/or lower PFAS plasma concentrations. For example, the NHANES (2003-2004) measured a median PFOA serum concentration of 3.8 ng/mL and the Canadian Health Measures Survey (2007-2009) measured a median PFOA concentration of 2.46 ng/mL. The study by Fassler et al., examining children in the Cincinnati area, had the highest media PFOA concentration of 7.7 ng/mL, which could be a reason for their borderline positive associations. Of the studies demonstrating significant positive correlations, cohorts were often at higher risk, including the elderly or obese children, or had a higher median PFOA concentration from 4.6-14.82 ng/mL.



Key physiological endpoints from studies that show non-significant or inverse associations between PFAS exposure and markers of T2D risk.

Citation	Cohort	PFAS Serum Concentrations	Physiological Endpoints
Liu et al., 2018 ( <a href="#">48</a> )	NHANES (2013-2014)	Geometric mean PFOA (ng/mL): <ul style="list-style-type: none"> <li>Total PFOA = 1.86</li> <li>n-PFOA = 1.75</li> <li>br-PFOA = 0.08</li> </ul> Geometric mean PFOS (ng/mL): <ul style="list-style-type: none"> <li>Total PFOS = 5.28</li> <li>n-PFOS = 3.70</li> <li>br-PFOS = 1.39</li> </ul>	↓Decreased HbA1c significantly associated with PFOA No significant associations between PFAS examined and 2hr glucose (GTT), insulin levels, or HOMA-IR
Nelson et al., 2010 ( <a href="#">49</a> )	NHANES (2003-2004)	Median (ng/mL): <ul style="list-style-type: none"> <li>PFOA = 3.9</li> <li>PFOS = 21.0</li> <li>PFHxS = 1.8</li> </ul>	No significant associations between PFAS examined and HOMA-IR.
Fisher et al., 2013 ( <a href="#">50</a> )	Canadian Health Measures Survey (2007-2009)	Geometric mean PFOA (ng/mL): <ul style="list-style-type: none"> <li>Male = 2.92</li> <li>Female = 2.18</li> </ul> Geometric mean PFOS (ng/mL): <ul style="list-style-type: none"> <li>Male = 11.0</li> <li>Female = 6.99</li> </ul>	No significant associations between PFAS examined and fasting insulin, fasting glucose, and HOMA-IR.
Fleisch, 2017 ( <a href="#">51</a> )	Mother-child pairs in Project Viva (mothers recruited 1999-2002, mid-childhood samples from 2007-2010)	Mid-childhood PFAS concentrations (Geometric mean, ng/mL): <ul style="list-style-type: none"> <li>PFOA: 4.2</li> <li>PFOS: 6.2</li> <li>PFDeA: 0.3</li> </ul>	↓Decreased HOMA-IR significantly associated with mid-childhood concentrations of PFOA, PFOS, and PFDeA Sex-stratified analysis showed PFOA, PFOS, and PFDeA concentrations inversely associated with HOMA-IR in females only
Fassler et al., 2019 ( <a href="#">52</a> )	Participants were 6 to 8-year old girls from Greater Cincinnati (2004-2006)	Median (ng/mL): <ul style="list-style-type: none"> <li>PFOA: 7.30</li> <li>PFOS: 18.60</li> <li>PFDeA: 0.30</li> </ul>	Borderline inverse association between insulin resistance and PFOA, PFOS, and PFDeA

Down arrow = decreased.

Furthermore, there have also been many studies performed investigating the association between exposure to PFAS and incidence of T2D, although associations often varied depending on the specific PFAS examined. In 2009, Lundin et al. published data analyzing a highly exposed cohort of employees from a 3M company in Cottage Grove, Minnesota that produced PFOA employed between 1947 and 1997 (53). The authors reported an increased risk of diabetes and diabetes-related deaths in workers with moderate exposure (HR = 3.7; 95% CI: 1.4-10.1), but not in low or high exposures. A 2018 study by Sun et al. examined the association between PFAS exposure and T2D in the Nurses' Health Study II (NHSII), a prospective cohort of female nurses in the U.S. enrolled who provided blood samples in 1995-2000 (54). Higher plasma concentrations of PFOS and PFOA were positively associated with increased risk of T2D. For PFOA and PFOS, respectively, comparing first and third tertiles,  $OR_{PFOS} = 1.62$  (95% CI: 1.09, 2.41;  $p_{trend} = 0.02$ ) and  $OR_{PFOA} = 1.54$  (95% CI: 1.04, 2.28;  $p_{trend} = 0.03$ ). However, plasma concentrations of PFHxS, PFNA, and PFDA did not show association with T2D. As mentioned previously, Zeeshan et al. examined data from Chinese adults from the "Isomers of C8 Health Project in China (46)." In addition to finding positive associations between PFAS and several diabetes indicators, they also found that serum concentrations of all PFAS were associated with increased risk of T2D in adults. Branched PFOS had the highest risk (OR=5.41, 95% CI: 3.68-7.96). The T2D incidence associations were generally stronger in women compared to men. In 2018, He et al. examined cross-sectional data from NHANES (2003-12) in order to examine the association of serum PFAS levels with diabetes prevalence (55). It was found that PFOA levels were strongly associated with diabetes prevalence in men (OR: 2.66, 95% CI: 1.63-4.35;  $p$  for trend = 0.001), but not women. Furthermore, PFOS, PFHxS, and PFNA were not found to be associated with diabetes prevalence regardless of gender. In 2019, Girardi et al. published an analysis of the association of PFAS exposure and mortality risk in a cohort of workers for a factory that produced PFOA and PFOS compared to workers from a metalworking factory from 1960-2008 (56). The authors found an increased relative risk for mortality from diabetes consequences (RR: 5.95; CI: 1.08-32.8) from workers in the PFAS factory compared to the metalworking factory. In 2017, as mentioned previously, Cardenas et al. examined participants of the Diabetes Prevention Program (DPP) (43). Although the authors found several PFAS associated with increased insulin resistance (HOMA-IR) and HbA1c levels, they did not find plasma PFAS concentrations to be strongly associated with incidence of T2D. However, the same group performed another study in 2019, including results from the Diabetes Prevention Program Outcomes Study (DPPOS), which was a 15-year follow-up study from the DPP (57). A doubling of PFOA was positively associated with a 14% increase in diabetes risk (HR 1.14, 95% CI 1.04, 1.25). However, this association was not observed in the lifestyle intervention group.

Although several studies demonstrated a positive association between exposure to PFAS and risk of T2D, various studies have also determined there to be a non-significant or inverse association between exposure to PFAS and risk of T2D. In 2019, Donat-Vargas et al. examined participants from the Swedish Vasterbotten Intervention Programme (VIP) prospective cohort, a sub-cohort of the Northern Sweden Health and Disease Study begun in 1985 (58). When comparing the first and third tertile, the odds ratio of T2D for the sum of PFAS was 0.52 (95% CI: 0.20, 1.36). Overall, PFAS was determined to be weakly inversely associated with the risk of future T2D, but mostly non-significant. Multiple groups have also analyzed the effects of PFAS exposure on T2D risk in an



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significant association between PFOA exposure and incidence of T2D ([59](#), [60](#)) or that PFAS serum levels were negatively associated with T2D: PFHxS (OR (CI))= 0.74 (0.71–0.77), PFOA (OR (CI))= 0.87 (0.89–0.91), PFOS (OR (CI))= 0.86 (0.82–0.90), and PFNA (OR (CI))= 0.94 (0.88–1.00) ([61](#)). However, these studies are limited by the cross-sectional nature and the fact that validation of T2D was only done for those who self-reported diabetes. In 2020, a study by Charles et al. analyzed samples from women in the Norwegian Women and Cancer study ([62](#)). Repeated measures of blood samples were collected prior to T2D diagnosis (2001/02) and then after T2D diagnosis (2005/6) as well as from non-diabetic controls. This study found no significantly increased OR for T2D incidence or prevalence with PFAS exposure. A recent study by Schilleman et al. published in 2021 examined how PFAS-related metabolites associate with risk of T2D in the Swedish VIP prospective cohort ([63](#)). PFAS levels were found to correlate positively with both glycerophospholipids and diacylglycerols. However, glycerophospholipids were found to associate inversely with risk of T2D, while diacylglycerols correlated positively with risk of T2D. These findings may play a role in understanding the conflicting results found for the association of T2D and PFAS levels in some epidemiological studies. In 2021, Han. et al. examined participants in a case-control study from Shandong Province, East China, recruited from 2016-2017 ([64](#)). The authors found that exposure to PFOA, PFOS, PFDA, PFUnDA, PFHxS, and 6:2 Cl-PFESA were all inversely associated with risk of T2D with significant ORs <1. However, serum PFOA was positively associated with fasting plasma glucose levels ( $\beta = 0.04$ , 95% CI: 0.0003, 0.08), which may promote T2D.

Although many studies have shown positive associations between PFAS exposure and T2D risk, others have shown inverse or no associations which may be due to exposure levels in a particular population or non-monotonic dose responses ([65](#), [66](#)). In 2018, Mancini and colleagues analyzed the link between T2D and dietary exposure to PFOA and PFOS in the E3N (Etude Epidémiologique auprès de femmes de l'Education Nationale) prospective cohort, which includes women in France between 40-60 years old and 15 years of follow-up ([65](#)). The mean dietary exposure was estimated to be 0.50ng/kg/day PFOS and 0.86ng/kg/day PFOA. This study found PFOA to have an inverse U-shape association with T2D, with the strongest positive effect on non-obese women. PFOS had a positive nonlinear association with T2D only among the non-obese women. In 2021, Duan et al. investigated the associations of various PFASs to T2D risk over a range of exposure levels ([66](#)). They examined data from participants in a case-control study, with half having confirmed T2D and half controls. They found that serum PFHxS and PFHpA displayed an inverted U-shaped dose-response, highest significant positive association with T2D risk in the middle tertiles. However, most other serum PFAS demonstrated a negative association with T2D risk, especially at higher levels.

## Future directions

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Analysis of PFAS in epidemiological studies have uncovered important associations between single PFAS and diabetes risk factors and incidence, and it will continue to be useful for emerging PFAS. However, this approach lacks the environmental relevance that comes with evaluating more realistic environmental mixtures. It will be important in the future to advance these findings by examining PFAS mixtures and incorporating the use of mixture statistics. Mixtures can interact to produce combination effects other than those predicted by individual constituent chemicals, including in-

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couple recent studies have begun to employ mixture statistics in their approach to estimating the combined effects of PFAS mixtures on diabetes incidence and estimated glomerular filtration rates. In 2021, Lin et al. published a follow-up study examining participants in the Diabetes Prevention Program trial (DPP, 1996–2002) and the later Outcomes Study (DPPOS, 2002–2014) (67). The authors evaluated the associations between PFAS levels and repeated measures of eGFR, individually, for PFOA, PFOS, PFHxS, EtFOSAA, MeFOSAA, and PFNA. It was found that, individually, total PFOS was observed to be inversely associated with eGFR: each doubling of total plasma PFOS was associated with  $-1.50 \text{ mL/min/1.73 m}^2$  lower eGFR (95% CI:  $-2.81, -0.19$ ). Quantile-based g-computation was then used to estimate the effects of the six PFAS as a mixture. As a mixture, it was found that each quartile increase in concentration of the PFAS mixture was associated with  $-2.26 \text{ mL/min/1.73 m}^2$  lower eGFR (95% CI:  $-4.12, -0.39$ ) at DPPOS Year 5, ~9 years since the PFAS measurements. Furthermore, in 2022, Park et al. published a study examining the association of PFAS with incident diabetes in the Study of Women's Health Across the Nation Multi-Pollutant Study (SWAN-MPS) (1999–2017) (68). The authors evaluated the associations between PFAS levels and diabetes incidence, individually, and found that the HR (95% CI) comparing the first and third tertile was 1.67 (1.21, 2.31) for n-PFOA ( $p_{\text{trend}} = 0.001$ ), 1.58 (1.13, 2.21) for PFHxS ( $p_{\text{trend}} = 0.003$ ), 1.36 (0.97, 1.90) for Sm-PFOS ( $p_{\text{trend}} = 0.05$ ), and 1.85 (1.28, 2.67) for MeFOSAA ( $p_{\text{trend}} = 0.0004$ ). Quantile-based g-computation was then used to evaluate the combined effects of seven PFAS as a mixture (n-PFOA, PFNA, PFHxS, n-PFOS, Sm-PFOS, MeFOSAA, and EtFOSAA). Exposure to the PFAS mixture was associated with an HR of 2.62 (95% CI 1.12, 6.20). As PFAS exist as complex mixtures in real-life settings, it will be important in the future to consider the impacts of PFAS as a complex mixture to our health. Furthermore, although in this review we have explored the effects of PFAS on T2D incidence and risk factors, the mechanisms potentially involved in these effects remain largely unknown.

Animal models have been used as a method for uncovering underlying mechanisms related to the effects of PFAS on glucose and insulin metabolism, as outlined in Table 3. A couple studies performed in rodent models demonstrated contrasting conclusions. In both, mice were exposed to PFOA at a dose of 1.25 mg/kg/day for 28 days. In 2015, Yan et al., PFOA exposure in mice was associated with increased insulin sensitivity and increased activation of the PI3K-AKT signaling, which is downstream of insulin receptor signaling (69). In 2017, Zheng et al. also exposed mice to PFOA for 28 days and observed that PFOA exposure increased blood glucose and decreased glycogen and glucose in the liver, while insulin levels in the blood were not significantly affected (70). Recently, *in vitro* studies have also been used to investigate mechanisms related to the effects of PFAS on glucose and insulin metabolism and  $\beta$ -cell function (73). In addition to their rodent model, Qin et al. exposed pancreatic  $\beta$ -cells to PFOS, which decreased their insulin release capacity following glucose stimulation (71). When the authors also exposed mice to PFOS for 28 days, pancreas weight and islet size were decreased. Duan et al. also demonstrated that exposure of pancreatic  $\beta$ -cells to PFOS resulted in impaired insulin secretion, possibly through inhibition of SIRT1 activity and decreased mitochondrial ATP production and  $\text{Ca}^{2+}$  influx (72). Further studies examining the effects of PFAS exposure on  $\beta$ -cell function and modulation of signaling pathways are needed in order to gain a better understanding of the mechanisms linking PFAS exposure to diabetes.

Key mechanistic endpoints from animal and *in vitro* PFAS exposure studies related to glucose and insulin metabolism.

Citation	Experimental Model	PFAS Serum Concentrations or <i>in vitro</i> Exposures	Physiological Endpoints
Yan et al., 2015 ( <a href="#">69</a> )	Male Balb/c mice (age 6-8 weeks) were administered PFOA diluted in water for 28 days at the following doses: 0, 0.08, 0.31, 1.25, 5, and 20 mg/kg/d <i>via</i> oral gavage.	Mean Serum PFOA: <ul style="list-style-type: none"> <li>0.08 mg/kg/d – 2.24 ug/mL</li> <li>0.31 mg/kg/d – 8 ug/mL</li> <li>1.25 mg/kg/d – 25 ug/mL</li> <li>5 mg/kg/d – 55 ug/mL</li> <li>20 mg/kg/d – 105.3 ug/mL</li> </ul>	<p>↑ Increased insulin and glucose sensitivity (GTT and ITT tests).</p> <p>↑ Increased activation of PI3K-AKT signaling pathway.</p>
Zheng et al., 2017 ( <a href="#">70</a> )	Male Balb/c mice (age 6-8 weeks) were administered 1.25 mg/kg/d PFOA for 28 days.	Mean PFOA (± SE): <ul style="list-style-type: none"> <li>55.5 ± 0.50 µg/mL (serum)</li> <li>125.9 ± 10.0 µg/g (liver)</li> </ul>	<p>↑ Increased fasting blood glucose</p> <p>↓ Decreased glycogen and glucose content in the liver.</p> <p>No significant change in blood insulin and insulin sensitivity</p>
Qin et al., 2022 ( <a href="#">71</a> )	<i>In vivo</i> : C57BL/6 mice administered 1-10 mg/kg/day PFOS <i>via</i> oral gavage for 28 days.  <i>In vitro</i> : Mouse β-TC-6 pancreatic cells	Mean Serum PFOS in mice (± SE): <ul style="list-style-type: none"> <li>0 mg/kg/d – 3 ± 2 µmol/L</li> <li>1 mg/kg/d – 63 ± 9 µmol/L</li> <li>5 mg/kg/d – 234 ± 27 µmol/L</li> <li>10 mg/kg/d – 348 ± 47 µmol/L</li> </ul> β-TC-6 cells: 48-hour exposure to PFOS at 50-100 µmol/L	<p>↓ PFOS exposure in male mice caused reductions of pancreas weight and islet size.</p> <p>↓ PFOS exposure decreased the insulin release capacity of β-TC-6 cells after glucose stimulation.</p>
Duan et al., 2021( <a href="#">72</a> )	Mouse β-TC-6 pancreatic cells.	β-TC-6 cells: exposure to PFOS at 0-200 µM for 24 or 48 hours.	<p>↓ PFOS exposure (48 h) impaired glucose stimulated insulin secretion.</p> <p>PFOS exposure decreased mitochondrial glucose-induced ATP production and Ca<sup>2+</sup> influx.</p>

Down arrow = decreased.

Up arrow = increased.

Overall, while studies have long been shown to demonstrate the major impact of genetic and life-style factors on the development and progression of T2D ([10–12](#)), emerging studies examined in this review have begun to investigate the impact of exposure to PFAS on glucose and insulin homeostasis as well as T2D incidence. In general, there seems to be strong evidence supporting a link between PFAS exposure and increased glucose levels and insulin resistance in individuals with multiple risk factors for diabetes, while PFAS exposures found in national surveys do not demonstrate consistent positive associations. In most cases however, these associations have been shown to vary depending on the specific PFAS examined and the gender of the subject.

The association of PFAS exposure with T2D incidence remains convoluted. It is possible these studies have produced such varied results due to an inverse U-shaped association between PFAS exposure and T2D incidence, as a couple studies have demonstrated ([67](#), [68](#)). Continued investigations into the relationships between PFAS exposure and T2D need to be done in order to better understand the impacts of environmental pollutants on diabetes progression and prevalence.

## Author contributions

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Conceptualization, MP; Writing—Original Draft Preparation, MP and KR; Writing—Review & Editing, MP and KR. All authors have read and agreed to the submitted version of the manuscript.

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## Conflict of interest

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The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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